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4 **Guideline on assessing the toxicological risk to human**  
5 **health and groundwater communities from veterinary**  
6 **pharmaceuticals in groundwater**  
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## 33 **Executive summary**

34 Groundwater is a source of human drinking water and also provides a very unique ecosystem with  
35 vulnerable aquatic communities. This guideline provides a methodology for performing a risk  
36 assessment of groundwater for human health and aquatic ecosystems for veterinary medicinal  
37 products. A risk assessment needs to be performed when the concentration of a VMP in groundwater is  
38 equal to, or above 0.1 µg/l. In addition, this guideline specifies cases where a risk assessment may be  
39 needed for highly toxic or persistent substances, when the predicted concentration in groundwater is  
40 lower than 0.1 µg/l. The guideline complements existing guidelines such as the CVMP guideline on  
41 environmental impact assessment for veterinary medicinal products in support of the VICH guidelines  
42 GL 6 and GL 38. These guidelines provide a methodology for risk assessment of veterinary medicines,  
43 but for groundwater they only provide a methodology for exposure assessment.

## 44 **1. Introduction and legal background**

45 The need for an assessment of veterinary medicinal products (VMPs) in groundwater is included in the  
46 VICH GL 38 "Environmental Impact Assessment for Veterinary Medicinal Products Phase II" (VICH  
47 2005) and the CVMP guideline "*Environmental Impact Assessment for Veterinary Medicinal Products in  
48 support of the VICH guidelines GL6 and GL38*" (EMA, 2016), referred to as the CVMP Technical  
49 Guidance Document (TGD). This guideline gives further technical support to the implementation of the  
50 VICH guidelines GL6 (VICH 2000) and GL38 (VICH 2005) on the environmental risk assessment (ERA)  
51 of VMPs. As part of the ERA, the predicted environmental concentration (PEC) in groundwater of a  
52 veterinary medicine needs to be compared against the value of 0.1 µg/l. The value of 0.1 µg/l is the  
53 groundwater quality standard (GQS) for pesticides and biocides according to the Groundwater Directive  
54 2006/118/EC (GWD), Annex I. Up to now, and based on the recommendations in the CVMP TGD (EMA,  
55 2016) the CVMP has followed the approach below:

- 56 • Concentrations above 0.1 µg/l have been considered unacceptable for all substances, regardless of  
57 their intrinsic hazardous properties. In this situation, applicants could refine the PEC groundwater  
58 with additional data (e.g., modelling, more studies e.g. on degradation in manure, mitigation  
59 measures), or, if this was not possible, restrictions to the proposed uses were required to avoid  
60 predicted values equal to or above 0.1 µg/l.
- 61 • For concentrations below 0.1 µg/l, no risk was anticipated and no further regulatory action was  
62 required.

63 When this value of 0.1 µg/l was introduced into the European groundwater legislation as a GQS, it was  
64 established as a precautionary limit value as this was the usual limit of detection of these kinds of  
65 substances in water (with concentrations below this limit being difficult to quantify due to analytical  
66 limitations). Consequently, predicted concentrations below 0.1 µg/l were interpreted as equivalent to  
67 'no emission into groundwater', thus concluding that any obligation to prevent entry of hazardous  
68 substances into groundwater was fulfilled. However, analytical methods are continuously improving  
69 with the result that it is now often possible to quantify substances below the current limit value of 0.1  
70 µg/l. Thus, the limit value is no longer guided by the limit of detection, and in certain situations  
71 additional considerations might be appropriate for substances for which concentrations have been  
72 detected or predicted below this value. This is especially the case when their safe concentration in  
73 water is significantly lower than this limit value of 0.1 µg/l.

74 It is also important to consider that the current Article 6(1) of the GWD mandates that all Member  
75 States shall take measures to prevent the input into the groundwater compartment of substances  
76 listed in points 1 to 6 of Annex VIII of the Water Framework directive (2000/60/EC) (so called, 'zero-  
77 tolerance substances'). Annex I of the GWD also lays down GQS for entry into groundwater of plant  
78 protection products, biocides and their metabolites. This Directive, however, does not contain any  
79 specific provisions for VMPs nor for those active ingredients in VMPs that are also active ingredients in  
80 pesticides or biocides. For example, certain VMPs contain the same substances as pesticides or  
81 biocides or meet the definitions of hazardous substances according to Annex VIII of Directive  
82 2000/60/EC, e.g. organohalogens.

83 The Phase II ERA guideline for VMPs (VICH, 2005) notes that 'groundwater is a natural resource and  
84 should not only be assessed with regards to public health but also to possible harmful effects to the  
85 biota of groundwater'. The guideline, however, does not provide any further information on how to  
86 assess possible harmful effects to human health or ecosystems, i.e., when the predicted concentration  
87 in groundwater is estimated to be  $\geq 0.1 \mu\text{g/l}$  and for situations where the predicted safe concentrations  
88 (i.e., predicted no effect concentration (PNEC)) for the aquatic compartment is below the value of  
89  $0.1 \mu\text{g/l}$ . For certain VMPs the PNECs are below  $0.1 \mu\text{g/l}$  and a review of environmental quality  
90 standards (EQS) for surface waters (values similar to PNEC values) has revealed that the safe annual  
91 average concentration could be three or more orders of magnitude lower than the GQS of  $0.1 \mu\text{g/l}$  (UK  
92 TAG, 2012). For example, the EQS for annual average concentrations of the substance cypermethrin in  
93 surface waters is  $8 \times 10^{-5} \mu\text{g/l}$  (Directive 2013/39/EU as regards priority substances in the field of water  
94 policy). The taxa that are specifically sensitive to these kinds of compounds are also present in  
95 groundwater. Thus, this example shows that a universal groundwater quality standard of  $0.1 \mu\text{g/l}$  may  
96 not be viewed as a safe value for all surface water and groundwater species.

## 97 **2. Scope**

98 This guideline outlines the methodology for performing a human health and environmental risk  
99 assessment for groundwater, considering:

- 100 1. The environmental and human health risk assessment approach for the groundwater compartment  
101 for substances with a predicted environmental concentration (PEC) above the GQS of  $0.1 \mu\text{g/l}$ .
- 102 2. An environmental risk assessment for a substance with PEC in groundwater below  $0.1 \mu\text{g/l}$  when  
103 scientific evidence indicates that concentrations below  $0.1 \mu\text{g/l}$  might pose a risk to groundwater  
104 ecosystems (for instance, if after a Phase II assessment, based on the aquatic PNEC it can be  
105 foreseen that the  $\text{PNEC}_{\text{groundwater}}$  will be lower than  $0.1 \mu\text{g/l}$ ).

106 When determining the potential impact of a VMP on groundwater, consideration has to also be given by  
107 the applicant to relevant European Union legislation on groundwater. In view of this, the CVMP  
108 considers that substances that come within the scope of points 1 to 6 of Annex VIII to the WFD, and  
109 active substances of VMPs that are also used as pesticides and/or biocides should be subject to the  
110 same limits, as laid down in Annex II of the GWD, i.e. that they should not enter groundwater at  
111 concentrations  $\geq 0.1 \mu\text{g/l}$ .

112 **3. Quantification of risk to humans from residues in**  
 113 **groundwater**

114 The risk to humans of VMPs in groundwater is associated with the use of groundwater as drinking  
 115 water. The risk assessment in this guideline is therefore based on scenarios following human  
 116 consumption of drinking water derived from groundwater.

117 In order to assess the risk of groundwater contamination by VMPs for human health a maximum  
 118 tolerable concentration in drinking water ( $MTC_{dw}$ ) ( $\mu\text{g/l}$ ) needs to be calculated following the same  
 119 methodology used by the World Health Organization (WHO, 2011).

120 
$$MTC_{dw} = \frac{ADI \times BW \times P}{C_w} \quad (\text{Equation 1})$$

121 The recommended default values to be used in the calculation of the  $MTC_{dw}$  and their units are listed in  
 122 Table 1.

123 **Table 1.** Recommended default values to be used in the calculation of the  $MTC_{dw}$  in Equation 1 and  
 124 their units.

	Parameter description	Unit	Default value
BW	Body weight	[kg]	60
$MTC_{dw}$	Maximum tolerable concentration in drinking water	$[\mu\text{g.l}^{-1}]$	–
ADI	Acceptable daily intake	$[\mu\text{g.kg}^{-1}.\text{day}^{-1}]$	–
P	Fraction of the ADI to be used by the intake of drinking water	[–]	0.1*
$C_w$	Daily consumption of drinking water	$[\text{l.d}^{-1}]$	2

125 \* An indicative value. See text for further details

126 The default values of 60 kg for the weight of the human body and 2 litres for the daily consumption of  
 127 drinking water are based on the recommendations made by WHO and used within the WHO  
 128 methodology. Any deviation from the default values in Table 1 needs to be accompanied by a solid  
 129 scientific justification.

130 The CVMP establishes acceptable daily intake (ADI) values for all active ingredients for use in VMPs for  
 131 administration to food producing animals. The ADI is a toxicological risk limit, which is identical to the  
 132 Tolerable Daily Intake (TDI) used within, for example, the WHO terminology. It represents a  
 133 toxicologically safe value even if exposure is prolonged. It is recommended to operate with 10% as an  
 134 indicative maximum percentage of the ADI to be used by drinking water. However, in principle the  
 135 exact fraction needs to be defined on a case by case basis as the acceptable daily intake via drinking  
 136 water depends on the anticipated daily intake via other sources and hence by the established MRLs,  
 137 meaning that the actual P value may in practice be higher or lower than 0.1. It is of paramount  
 138 importance that the total intake of the VMP by consumers (via water and food products) does not  
 139 exceed the ADI.

140 Based on the information on the predicted concentration in groundwater ( $PEC_{gw}$ ) and the maximum  
141 tolerable concentration in drinking water ( $MTC_{dw}$ ), the potential risk of the VMP to humans via drinking  
142 water can be assessed by the risk quotient in Equation 2 below.

$$143 \quad RQ_{gw} = \frac{PEC_{gw}}{MTC_{dw}} \text{ (Equation 2)}$$

144 In cases where the  $RQ_{gw}$  exceeds 1.0 a potential risk for groundwater has been identified and a risk  
145 refinement needs to be done. This may include a refinement of the  $PEC_{gw}$  and/or a scientific  
146 justification for deviating from the default input parameters in the  $MTC_{gw}$  calculation. For a refinement  
147 and final establishment of the  $PEC_{gw}$  the  $K_{oc}$  and  $DT_{50}$  are typically needed in order to run the  
148 appropriate models, which are available in the ERA according to the VICH GLs.

149 If the  $RQ_{gw}$  is below 1.0, the risk to human health is considered acceptable.

## 150 **4. Quantification of risk to groundwater ecosystems**

151 In addition to the human health assessment, an environmental risk assessment for groundwater  
152 ecosystems is also needed. Contamination of groundwater may permanently eradicate entire unique  
153 groundwater communities due to the low or absent ability to re-colonise any affected habitats. In  
154 addition, because of the absence of primary producers in groundwater systems, the self-purification  
155 process may be disturbed and the original state cannot easily be restored.

### 156 **4.1. Definition of groundwater ecosystems of concern**

157 In the environmental risk assessment procedure according to the VICH guidelines GL6 (VICH 2000)  
158 and GL 38 (VICH, 2005), ecosystems in the aquatic compartment are defined by a few representative  
159 habitats e.g. a stream, a pond, freshwater sediment for freshwater surface ecosystems or pelagic  
160 water, and marine sediments for marine ecosystems. Probably the most representative groundwater  
161 habitats are on hypogean karst (fractures, channels, caves), and alluvial gravel interstitial systems. For  
162 the protection of the whole groundwater compartment, it is also necessary to protect spring water.  
163 Springs are an ecotone, a transition between groundwater and surface water bodies, and in this  
164 context are also considered groundwater, as they may fill gravel pits and ponds or lakes without other  
165 influx from surface water.

### 166 **4.2. Environmental risk assessment considerations**

167 The assessment of the environmental risks of VMPs is framed by internationally agreed guidelines (i.e.,  
168 VICH guidelines GL6 and GL38), which prescribe the use of data from a limited number of tests on  
169 fate, behaviour and ecotoxicity of the VMP. The current guidelines do not include ecotoxicity testing for  
170 the determination of the safe concentration of a substance in water with species other than surface  
171 water species (i.e., algae, daphnids and fish species). Experimental data obtained from these three  
172 groups of organisms may be extrapolated to predict safe concentrations for highly adapted  
173 groundwater biota. However, this extrapolation is only possible when the unique biotic and abiotic  
174 conditions of groundwater ecosystems are taken into account. Any additional uncertainties should be  
175 addressed through the use of more adequate assessment factors.

176 The groundwater ecosystem is considered more vulnerable than many other aquatic ecosystems. The  
177 concept of ecosystem vulnerability, defined as the "exposure to contingencies and stress, and the

178 difficulty in coping with them" (Millennium Ecosystem Assessment report, 2005), comprises three  
179 major dimensions:

180 *i) Exposure to stress, perturbations and shocks;*

181 *ii) Sensitivity of exposed species to the stress or perturbation, including their capacity to*  
182 *anticipate and cope with the stress;*

183 *iii) Resilience of the exposed ecosystems, and species in terms of their capacity to absorb shocks*  
184 *and perturbations while maintaining function.*

185 The 3 bullets mentioned above have been considered when determining the most appropriate  
186 assessment factors when extrapolating from data on surface water species.

#### 187 **4.2.1. Exposure to stress, perturbations, and shocks:**

188 An important element to be considered for the ERA of groundwater ecosystems is the duration of  
189 exposure of chemicals to any resident communities. There are indications that groundwater  
190 ecosystems have a more prolonged period of exposure time, given that this ecosystem is generally  
191 colder than surface water systems, and this aspect combined with a complete absence of sunlight (at  
192 least in the vast majority of cases), could result in decreased biotic and abiotic degradation processes  
193 (Yagi et al., 2010; Bulog and Mali Bizjak, 2014) leading to chemical concentrations persisting longer.  
194 In most surface water systems (e.g. rivers), the residence time of water (the average amount of time  
195 that a molecule of water spends in a particular system) ranges from a few days to a few weeks.  
196 However, in groundwater ecosystems such as porous aquifers, the residence time may be decades or  
197 even centuries (Freeze and Cherry, 1979). Consequently, whereas the baseline quality of the riverine  
198 surface water may recover after a relatively short period of time, the recovery in aquifers (if any) may  
199 require decades, or disturbances may even be irreversible.

#### 200 **4.2.2. Sensitivity of exposed species to the stress or perturbation,** 201 **including their capacity to anticipate and cope with the stress**

202 The sensitivity of groundwater species to an acute exposure of a chemical stressor may be equivalent  
203 or higher than that of a model aquatic species used for ecotoxicity testing (Blonk, 2014). Further, the  
204 chronic effects of chemical stressors to groundwater species are not known (Blonk, 2014; van Beelen,  
205 2007; Avramov, 2013).

206 Groundwater species have developed a number of metabolic adaptations for extreme energy saving to  
207 deal with starvation, for instance their metabolism is significantly lower than that of most other aquatic  
208 species. These adaptations can affect the species response to a chemical stressor, and make them  
209 more sensitive to the long term effects of chemical stressors than surface water species (Di Lorenzo et  
210 al., 2015). However, it is also important to note that the metabolic potential of groundwater species  
211 under sudden favourable conditions is higher for faster energy recovery and restoration of body  
212 reserves (Simčič et al., 2005; Hervant et al., 1997). Thus, the ability to respond to chemical stressors  
213 may be slower in groundwater species. The energetic cost of this response may cause a secondary  
214 effect on the survival and reproduction in exposed groundwater species in the long term, when  
215 compared to surface water species.

### 216 4.2.3. Resilience and complexity of groundwater ecosystems

217 Groundwater ecosystems are characterised by lower levels of complexity due to the allotrophy<sup>1</sup> and  
218 heterotrophy<sup>2</sup> of this system. This lower level of complexity also implies that re-colonisation after  
219 perturbation may be extremely slow, and that the restoration of an affected biotic community is  
220 unlikely (Culver and Pipan, 2009). Thus, groundwater ecosystems have low resilience to perturbations.  
221 Groundwater species are characterised by a high level of endemism, longevity, slow growth and low  
222 reproduction rates. The biological adaptations of these species contribute to the low plasticity of their  
223 community when reacting to repeated or chronic pressure due to chemical stressors.

224 Surface water ecosystems are characterised by a higher level of complexity, making them more able to  
225 recover from stress, mostly in the timeframe of one or few seasons.

226 Due to the low resilience to perturbations by groundwater ecosystems, recovery from the toxic effects  
227 of VMPs may not be possible within a realistic timeframe. Consequently, the permanent loss of  
228 endemic groundwater species is a realistic threat (Bulog and Mali Bizjak, 2014).

### 229 4.2.4. Conclusion on the use of assessment factors

230 The arguments presented above show that groundwater ecosystems may be more vulnerable than  
231 surface water ecosystems and may lack the ability to recover from perturbations. As a result, an  
232 additional assessment factor of 10 should be applied to extrapolate the  $PNEC_{\text{groundwater}}$  from the  
233  $PNEC_{\text{Surfacewater}}$ . A similar approach based on vulnerability, complexity of ecosystems and limited data  
234 availability is already applied when extrapolating the EQS for surface waters to an EQS for "other  
235 surface waters" (i.e., pelagic marine ecosystem) within the Water Framework Directive (Directive  
236 2013/39/EC).

$$237 PNEC_{\text{groundwater}} = PNEC_{\text{surfacewater}}/10$$

## 238 5. Risk assessment of groundwater ecosystems

### 239 5.1. Scenario 1: $PEC_{\text{groundwater}} \geq 0.1 \mu\text{g/l}$

#### 240 5.1.1. Exposure assessment

241 The initial concentration of a VMP in groundwater ( $PEC_{\text{groundwater}}$ ) is determined using the methods  
242 described in the CVMP TGD (2016). If the initial concentration exceeds the QOS of  $0.1 \mu\text{g/l}$ , the  
243 exposure can be refined. The  $PEC_{\text{surfacewater}}$  is first refined using the FOCUS<sup>3</sup> PEARL model (Focus,  
244 2000), using the parameters outlined in Appendix I of the CVMP TGD (2016). The exposure  
245 assessment can also be refined by determining degradation in manure in accordance with the  
246 respective CVMP guideline (EMA, 2011).

#### 247 5.1.2. Risk characterisation

248 In Phase II of the ERA the safe concentration for surface water species is determined ( $PNEC_{\text{surfacewater}}$ )  
249 based on experimental ecotoxicological data. The  $PNEC_{\text{surfacewater}}$  can be extrapolated to a

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<sup>1</sup> Referring to bioproduction being largely dependent on transport of resources from the surface

<sup>2</sup> Referring to the need to use organic carbon for growth.

<sup>3</sup> FOCUS: Forum for the Coordination of Pesticide Fate Models and Their Use



250 PNEC<sub>groundwater</sub> (see chapter 4). For this extrapolation, an additional assessment factor of 10 should be  
251 considered.

$$252 \quad \text{PNEC}_{\text{groundwater}} = \text{PNEC}_{\text{surfacewater}} / 10$$

253 Using the PNEC<sub>groundwater</sub>, the RQ for the groundwater compartment is determined.

$$254 \quad \text{RQ}_{\text{groundwater}} = \text{PEC}_{\text{groundwater}} / \text{PNEC}_{\text{groundwater}}$$

255 If the RQ<sub>groundwater</sub> is  $\geq 1$  based on this initial PEC<sub>groundwater</sub> the applicant can propose refinement of  
256 PEC<sub>groundwater</sub> values according to the supporting CVMP GL (EMA 2005). The PNEC<sub>groundwater</sub> can also be  
257 refined, using the standard Tier B surface water tests (but still including the additional assessment  
258 factor of 10). No further refinement of PNEC<sub>groundwater</sub> with experimental studies on groundwater  
259 species is possible, as long as no appropriate ecotoxicological long term test for groundwater species  
260 are available.

261 When a risk to the groundwater is indicated, the applicant should propose adequate risk mitigation  
262 measures to protect groundwater ecosystems. If no suitable risk mitigation measure/s can be applied,  
263 the risk for groundwater has to be addressed in the benefit risk assessment.

## 264 **5.2. Scenario 2: PEC<sub>groundwater</sub> < 0.1 µg/l and PNEC<sub>surfacewater</sub> ≤ 1** 265 **µg/l derived from Tier B surface water toxicity tests**

266 For most VMPs, the PEC<sub>groundwater</sub> will be  $\leq 0.1 \mu\text{g/l}$ . In these situations, a risk assessment for this  
267 compartment will not be needed, unless Tier B ecotoxicity data show that the PNEC<sub>surfacewater</sub> is  $\leq 1 \text{ g/l}$ .  
268 In this situation, the PNEC<sub>groundwater</sub> would be  $\leq 0.1 \mu\text{g/l}$  (i.e.,  $\text{PNEC}_{\text{groundwater}} = \text{PNEC}_{\text{surfacewater}} / 10$ ), and  
269 a risk to the groundwater compartment for a compound with a concentration  $< 0.1 \mu\text{g/l}$  may not be  
270 excluded. In this case, PNEC<sub>groundwater</sub> needs to be compared to the PEC<sub>groundwater</sub> in order to determine  
271 if there is a possible risk to the groundwater compartment, even if the PEC<sub>groundwater</sub> is  $\leq 0.1 \mu\text{g/l}$ .

272 When a risk to the groundwater is indicated, the applicant should propose adequate risk mitigation  
273 measures to protect groundwater ecosystems. If no suitable risk mitigation measure/s can be applied,  
274 the risk for groundwater has to be addressed in the benefit risk assessment.

## 275 **References**

276 Avramov, M. 2013. Sensitivity and stress of groundwater invertebrates to toxic pollution and changes  
277 in temperature (Doctoral dissertation). Technische Universität München. Retrieved from  
278 <https://push-zb.helmholtz-muenchen.de/deliver.php?id=5172>

279 Blonk, B. 2014. The effect of nutrients and toxic compounds on groundwater organisms in the Water  
280 Framework Directive (Master dissertation). Universiteit van Amsterdam.  
281 [http://ibed.uva.nl/research/research-groups/content/aquatic-environmental-](http://ibed.uva.nl/research/research-groups/content/aquatic-environmental-ecology/research/msc-projects/msc-projects.html)  
282 [ecology/research/msc-projects/msc-projects.html](http://ibed.uva.nl/research/research-groups/content/aquatic-environmental-ecology/research/msc-projects/msc-projects.html)

283 Bulog, B. and Mali Bizjak, L. Olm – cave salamander. In natural heritage of Bela krajina, Slovenia. Bela  
284 krajina Museum. 2014. pp 177-186. Retrived from [http://web.bf.uni-lj.si/bi/NATURA-](http://web.bf.uni-lj.si/bi/NATURA-SLOVENIAE/pdf/NatSlo_18_1_8.pdf)  
285 [SLOVENIAE/pdf/NatSlo\\_18\\_1\\_8.pdf](http://web.bf.uni-lj.si/bi/NATURA-SLOVENIAE/pdf/NatSlo_18_1_8.pdf)

286 Culver, D. C. and Pipan, T. The Biology of Caves and Other Subterranean Habitats. Publisher, Oxford  
287 University Press. 273 pp 2009. ISBN 0191551449.

- 288 Di Lorenzo, T., Di Marzio, W. D., Spigoli, D., Baratti, M., Messana, G., Cannicci, S. and Galassi, D. M. P.  
289 2015. Metabolic rates of a hypogean and an epigean species of copepod in an alluvial aquifer.  
290 *Freshw Biol*, 60, 426–435.
- 291 European Medicines Agency (EMA). 2016. Environmental Impact Assessment for Veterinary Medicinal  
292 Products in support of the VICH guidelines GL6 and GL38. EMEA/CVMP/ 418282/2005-Rev.1.  
293 [http://www.ema.europa.eu/docs/en\\_GB/document\\_library/Scientific\\_guideline/2009/10/WC5000](http://www.ema.europa.eu/docs/en_GB/document_library/Scientific_guideline/2009/10/WC5000)  
294 04389.pdf
- 295 European Medicines Agency (EMA). 2011. Guideline on determining the date of veterinary medicinal  
296 products in manure. EMA/CVMP/ERA/430327/2009.  
297 [http://www.ema.europa.eu/docs/en\\_GB/document\\_library/Scientific\\_guideline/2011/03/WC5001](http://www.ema.europa.eu/docs/en_GB/document_library/Scientific_guideline/2011/03/WC5001)  
298 04495.pdf
- 299 Hervant, F., Mathieu, J., Barré, H., Simon, K. and Pinon, C. 1997. Comparative study on the  
300 behavioral, ventilatory, and respiratory responses of hypogean and epigean crustaceans to long-  
301 term starvation and subsequent feeding. *Comparative Biochemistry and Physiology, Part A:*  
302 *Physiology*, 118:4, 1277–1283.
- 303 Freeze, R.A. and Cherry, J.A. 1979. *Groundwater*. Prentice-Hall, Englewood Cliffs, NJ, 604pp
- 304 FOCUS 2000. FOCUS groundwater scenarios in the EU review of active substances" v.2.2 - Report of  
305 the FOCUS Groundwater Scenarios Workgroup, EC Document Reference Sanco/321/2000 rev.2.  
306 [http://esdac.jrc.ec.europa.eu/public\\_path/projects\\_data/focus/gw/docs/FOCUS\\_GW\\_Report\\_Main](http://esdac.jrc.ec.europa.eu/public_path/projects_data/focus/gw/docs/FOCUS_GW_Report_Main)  
307 .pdf
- 308 Millennium Ecosystem Assessment. 2005. Island Press, Washington DC. Available online  
309 <http://www.MAweb.org.MA>, 2005
- 310 Official Journal of the European Union. Directive 2006/118/EC of the European Parliament and of the  
311 Council of 12 December 2006 on the protection of groundwater against pollution and  
312 deterioration [2006] OJ L327/19
- 313 Official Journal of the European Union. Directive 2013/39/EU of the European Parliament and of the  
314 Council of 12 August 2013 amending Directives 2000/60/EC and 2008/105/EC as regards priority  
315 substances in the field of water policy [2013] OJ L226/1.
- 316 Simič, T., Lukančič, S. and Brancelj, A. 2005. Comparative study of electron transport system activity  
317 and oxygen consumption of amphipods from caves and surface habitats. *Freshwater Biology*,  
318 50:3, 494–501.
- 319 UK Technical advisory group (UK TAG) - Proposals for further environmental quality standards for  
320 specific pollutants. Final report. 2012. Available online  
321 [https://www.wfduk.org/sites/default/files/Media/UKTAG%20Specific%20Pollutants\\_final\\_011412](https://www.wfduk.org/sites/default/files/Media/UKTAG%20Specific%20Pollutants_final_011412).  
322 pdf
- 323 van Beelen, P. 2007. Ecologische risicobeoordeling van grondwater. Rijksinstituut voor  
324 Volksgezondheid en Milieu (RIVM) no.711701055 , 44pp. Retrieved from  
325 <http://www.rivm.nl/bibliotheek/rapporten/711701055.html>

- 326 Veterinary International Conference on Harmonization (VICH). 2000. GL 6 - Guideline on  
327 environmental impact assessment (EIAS) for veterinary medicinal products – phase I.  
328 CVMP/VICH/592/98-FINAL.  
329 [http://www.ema.europa.eu/docs/en\\_GB/document\\_library/Scientific\\_guideline/2009/10/WC5000](http://www.ema.europa.eu/docs/en_GB/document_library/Scientific_guideline/2009/10/WC5000)  
330 04394.pdf
- 331 Veterinary International Conference on Harmonization (VICH). 2005. GL 38 - Guideline on  
332 environmental impact assessment for veterinary medicinal products phase II.  
333 CVMP/VICH/790/03-FINAL.  
334 [http://www.ema.europa.eu/docs/en\\_GB/document\\_library/Scientific\\_guideline/2009/10/WC5000](http://www.ema.europa.eu/docs/en_GB/document_library/Scientific_guideline/2009/10/WC5000)  
335 04393.pdf
- 336 World Health Organisation (WHO), 2011. Guidelines for Drinking-water Quality. Fourth Edition.  
337 Available online  
338 [http://www.who.int/water\\_sanitation\\_health/publications/2011/dwq\\_guidelines/en/](http://www.who.int/water_sanitation_health/publications/2011/dwq_guidelines/en/)
- 339 Yagi, J. M., Neuhauser, E. F., Ripp, J. a, Mauro, D. M. & Madsen, E. L. 2010. Subsurface ecosystem  
340 resilience: long-term attenuation of subsurface contaminants supports a dynamic microbial  
341 community. The ISME Journal, 4, 131–143.